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A review of microplastic impacts on seagrasses, epiphytes, and associated sediment communities

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1 **Microplastic impacts on seagrasses, epiphytes, and associated sediment communities**

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9 **Highlights**

- 10 • Microplastics accumulate within seagrass ecosystems, both on blades and in sediments.
- 11 • Potential microplastic impacts on seagrass ecosystem processes are examined.
- 12 • A special focus is given to plant, epiphyte, and sediment processes.
- 13 • Direction for future experimentation is provided.

14 **Abstract**

15 Microplastics have been discovered ubiquitously in marine environments. While their
16 accumulation is noted in seagrass meadows, little attention has yet been given to microplastic
17 impacts on valuable ecosystem functions. We initiate this discussion by synthesizing the
18 potential impacts microplastics may have on plant, epiphyte, and sediment processes. We
19 suggest that microplastics may harm epiphytes and seagrasses via impalement and light/gas
20 blockage, and increase local concentrations of toxins, causing a disruption in metabolic
21 processes. Further, microplastics may alter nutrient cycling by inhibiting dinitrogen fixation by
22 diazotrophs, preventing microbial processes, and reducing root nutrient uptake. They may also
23 harm benthic organisms via complications associated with ingestion and sediment characteristic

24 alteration. All impacts will be exacerbated by seagrasses' high microplastic trapping efficiency.
25 As microplastics become a permanent and increasing member of seagrass ecosystems it will be
26 pertinent to direct future research towards understanding the extent of microplastic impacts on
27 seagrass ecosystem processes.

28
29 **Keywords**

30 Seagrass; epiphytes; microplastics; photosynthesis; nutrient cycle; food web

31 **The State of Microplastics in Seagrass Ecosystems**

32 Microplastic debris, defined as plastic particles that measure <5mm in length and persist
33 in a variety of shapes and colors, have been increasing in global oceanic abundance. Inputs
34 include primary microplastics from commercially produced particles such as synthetic textile
35 fibers and cosmetic microbeads, as well as secondary microplastics that result from the
36 breakdown of larger plastics like water bottles and bags (Barnes et al., 2009; Rocha-Santos et al.,
37 2015; Shim et al., 2018). Microplastics are recalcitrant and remain in systems indefinitely
38 following their introduction (Bowley et al., 2021). They can be further subdivided into categories
39 based on their chemical composition (polyethylene, polyester, nylon etc.) and appearance (fiber,
40 bead, fragment) (Andrady 2011; Goss et al., 2018). Since their discovery, microplastics have
41 been found pervading many regions of the marine environment, including remote locations like
42 deep-sea sediments in the Arctic (Carpenter et al., 1972; Bergmann et al., 2017). Though marine
43 microplastics can be found ubiquitously, a large proportion of them accumulate on shorelines in
44 close proximity sources, making it pertinent to study the impacts of this pollution on populated
45 nearshore ecosystems (Thiel et al., 2013). Adverse effects of microplastics, such as
46 complications following ingestion by marine organisms and toxicity, have been reported in

47 numerous studies, further solidifying the need to advance investigation of this novel pollutant
48 (Nobre et al., 2015; Zhang et al., 2017; Rotjan et al., 2019; Dantas et al., 2020).

49 Coastal systems on every continent except Antarctica are colonized by a wide variety of
50 seagrass species, which are susceptible to many anthropogenic impacts due to their proximity to
51 developed areas (Short et al., 2007; Unsworth et al., 2018). Stressors such as disturbance from
52 trawling and dredging, aquaculture, and eutrophication have created a worldwide crisis in
53 seagrass communities (Duarte, 2002; Unsworth et al., 2018). The culmination of these threats has
54 resulted in a tenfold reduction in global seagrass ecosystems in the last forty years (Orth et al.,
55 2006). This decline is significant both economically and ecologically, as seagrasses provide
56 unequivocal services to the surrounding environment and human population (Nordlund et al.,
57 2018). Seagrass meadows are particularly valuable for their ability to sequester carbon at high
58 rates proportional to their biomass. Though they occupy only 0.1% of the total ocean floor,
59 seagrasses are responsible for 11%, or upwards of 83 million metric tons of organic carbon
60 sequestration every year, with a global storage capacity upwards of 19.9 Pg organic carbon
61 (Howard et al., 2017; Fourqurean et al., 2012). Other services include, but are not limited to,
62 preventing coastal erosion via wave impact reduction and sediment stabilization, improving
63 water quality, accreting sediments, and offering food and habitat to a variety of marine
64 organisms, many of which are commercially important or endangered (Cornell et al., 2018;
65 Fonesca et al., 1992; Nordlund et al., 2018; Nordlund et al., 2017).

66 Recently, microplastics have been discovered as yet another potential threat to seagrass
67 ecosystems. Seagrasses have been identified as substantial sinks for microplastics, both in the
68 sediment within meadows and adhered to seagrass blades (Goss et al., 2018; Jones et al., 2020).
69 Once present in the ecosystem they can become incorporated into food webs via blade

70 consumption and bioaccumulate in higher trophic levels (Goss et al., 2018). While impacts on
71 higher trophic organisms are important, little attention has yet been given to the effects of
72 microplastics on the ecosystem's foundation species: seagrasses themselves and their associated
73 epiphytic and benthic communities. In an effort to motivate and direct future research in this
74 area, as well as bring attention to the growing microplastic crisis in seagrass ecosystems, this
75 article synthesizes studies of known microplastic impacts on relevant algal, macrophytic, and
76 benthic communities and hypothesizes how they may impact seagrass ecosystem processes,
77 noting key gaps and opportunities in the literature.

78 **Seagrass Ecosystem Characteristics**

79 To understand the potential impact of microplastics on seagrasses, baseline ecosystem
80 characteristics must be established. Seagrasses are marine angiosperms that can form dense
81 meadows. They have extensive root and rhizome systems that stabilize sediments, while their
82 high above ground biomass and architectural complexity reduces water currents, causing
83 particulate matter to become trapped amongst blades and settle to the ground (Miyajima et al.,
84 2015; Smit et al., 2021). Because of this, seagrass beds accumulate large quantities of particulate
85 organic matter (POM), both onto their blades and to the sediments below (Ward et al. 1984). The
86 hydrodynamics of seagrass meadows also allow for the settlement of epiphytes; small sessile
87 organisms such as cyanobacteria, diatoms, crustose coralline algae, and macroalgae that adhere
88 to seagrass blades (Bologna et al. 1999). Epiphyte abundance can be influenced by nutrient
89 availability, water motion, light, and temperature, factors which similarly impact seagrass
90 abundance (Borum et al., 1985; Borowitzka et al., 2007). In some instances, these organisms can
91 account for more than half of a seagrass ecosystem's total aboveground primary production and
92 aboveground biomass (Morgan and Kitting, 1984; Pollard and Korure, 1993). They also provide

93 nutrients and spatial complexity, increasing faunal density in the ecosystem (Bologna et al.,
94 1999). However, they have adverse effects on seagrasses themselves, namely through the
95 reduction of light available for photosynthesis, reduction in the rate of CO₂, O₂ and nutrient
96 diffusion across the blade surface, and increase in drag (Nelson, 2017; Brodersen et al., 2015;
97 Brodersen et al., 2020).

98 Seagrass ecosystems sequester large quantities of carbon via photosynthesis, contributing
99 the most per unit area among all vegetated coastal habitats (IUCN, 2017). Carbon stored within
100 seagrass meadows is found in above and belowground plant biomass (Macreadie et al., 2014).
101 However, most of the stored carbon is found within sediments, stored via bacteria, algae,
102 detritus, and carbonates (Gacia et al., 2002). Because seagrass sediments are generally anoxic,
103 less than 10% of organic material entering the sediment decomposes, allowing for long term
104 storage (Gacia et al., 2002). The anaerobic rhizosphere of seagrass sediments also boasts
105 substantial contributions to the plant's nitrogen requirement, as bacterial nitrification and
106 denitrification processes interplay to regulate nutrient availability and seagrass primary
107 production (Capone & Taylor, 1980). Seagrass sediments are an important component of the
108 ecosystem from the perspective of long-term carbon storage and nutrient cycling.

109 Due to seagrasses' ability to reduce currents, house epiphytic communities, and trap and
110 accumulate particulate matter, they are poised as an ideal location for microplastic settlement.
111 Huang et al., (2021) found that seagrass ecosystems were enriched with microplastics 1.3 to 17.6
112 times compared to unvegetated sites, while Huang et al. (2020) found an enrichment factor of up
113 to 2.9. This significant accumulation, coupled with the ecosystems' diverse and important
114 inhabitants, positions seagrasses as a necessary system to consider microplastic impacts
115 (Nordlund et al., 2017). Further, seagrasses worldwide are highly valued for their unequivocal

116 ecosystem services, which are already under threat from numerous other stressors, making it
117 crucial to determine the relative threat of microplastics when assessing ecosystem health and
118 longevity, particularly because there is little to no widespread technology currently available to
119 remove microplastics from coastal systems (Duarte et al., 2002; Nordlund et al. 2018;
120 Malankowska et al., 2021). It is therefore vital to elucidate the potential impacts of microplastics
121 on this ecosystem and identify gaps in the literature to guide future experimentation regarding
122 this phenomenon.

123 **Microplastic Effects on Seagrass and Epiphyte Communities**

124 Once microplastics enter seagrass ecosystems, they can become incorporated into
125 epiphytic communities attached to seagrass blades. Microplastics in the water column get
126 trapped amongst seagrass plants due to reduced currents in a similar fashion as POM (De los
127 Santos et al., 2021). Then the rough surface of epiphytes provides a substrate onto which
128 microplastics can adhere and become trapped. Epiphytes overgrow the trapped microplastics,
129 keeping them attached to the blade surface (Fig. 1). This is supported by Goss et al. (2018) and
130 Datu et al. (2019), who both found that many microplastics on seagrass blades were embedded
131 within epiphyte assemblages. There is also emerging evidence that microplastic abundances are
132 significantly related to epiphyte density, suggesting that more epiphytes on a blade directly
133 correlate to more microplastics (Gerstenbacher et al., in prep). Some genres of seagrass, namely
134 *Posidonia*, are not only capable of trapping microplastics within their own ecosystem, but within
135 their exported ball shaped wrack, known as aegagropilae, which accumulate on beaches adjacent
136 to seagrass meadows, indicating that seagrasses may also play a role in exporting microplastics
137 out of marine environments, as well as trapping them within their own ecosystem (Sanchez-
138 Vidal et al., 2021).

139 While there are several documentations of microplastics embedded in epiphytic
140 assemblages on seagrass blades, there are no peer reviewed publications exploring the impacts of
141 this accumulation directly on epiphytes and seagrass blade surfaces (Table 1). Existing research
142 regarding the impacts of microplastics and related issues on marine algae including diatoms and
143 cyanobacteria, which are often part of epiphytic assemblages, as well as marine macrophytes,
144 may provide insight on potential impacts, directing future experimentation in the area.

145 One of the greatest threats associated with microplastics is their ability to absorb or
146 heteroaggregate with toxins including Persistent Organic Pollutants (POPs) and nanoparticles
147 (NPs). This occurs because many pollutants have a greater affinity for microplastics'
148 hydrophobic surface than seawater (Tueten et al., 2009; Wang et al., 2016). This affinity can
149 vary among different types of plastic, as physical and chemical attributes impact sorption
150 capabilities (Wang et al., 2016). More than 78% of these pollutants are considered toxic to
151 marine life, and their interactions with microplastics may increase their toxicity towards marine
152 algae (Rochman, 2013; Thiagarajan et al., 2019). The majority of contaminated microplastics
153 sink in nearshore marine systems where seagrass meadows reside (Bakir et al., 2014). Due to the
154 physical attributes of seagrasses, contaminated plastics are more likely to sink within seagrass
155 ecosystems than nearby bare sediments, likely coming in contact with epiphytic algae during the
156 process (de los Santos et al., 2021). Once trapped amongst an epiphytic community, POPs and
157 other toxins attached to microplastics can become bioavailable to algae, including those found in
158 epiphytic assemblages, via desorption (Heinrich & Braunbeck, 2019). This is evident in the
159 marine algae *Chlorella sp.*, which is capable of absorbing toxins transported by microplastics,
160 including nanoparticles and triphenyltin chloride (Thiagarajan et al., 2019; Thiagarajan et al.,
161 2021; Yi et al., 2020). This process is variable and depends on the type of plastic, its chemical

162 composition, and the type of algal species, however it is evident that microplastic associated
163 toxins are capable of being absorbed by marine algae (Heinrich & Braunbeck, 2019). Once
164 present in marine algae, these pollutants cause a variety of negative impacts. Generally, organic
165 pollutants can inhibit both light and dark reactions in photosynthesis, subsequently reducing
166 growth and vitality (Tomar et al., 2019). This impact is seen in higher order plants as well as
167 algae (Tomar et al., 2019). In marine algae, nanoparticles and other toxins result in oxidative
168 stress, a reduction in Chlorophyll A, cell lysis, and lipid membrane damage (Thiagarajan et al.,
169 2019; Thiagarajan et al., 2021; Yi et al., 2020). Cell membrane damage is particularly
170 concerning, as it may facilitate the internalization of further toxins (Thiagarajan et al., 2021).
171 While some level of pollutant contamination already exists in seagrass ecosystems,
172 microplastic's role as a vector brings these toxins directly to epiphytic communities, potentially
173 making them more locally available within blade microenvironments, exacerbating negative
174 effects (Huang et al., 2020., McLachlan et al., 2001).

175 Beyond the vectoring of pollutants to epiphytic communities, microplastics themselves
176 raise a threat to these assemblages. Contact between marine algae and plastic polymers,
177 including the common Polyvinyl Chloride (PVC) and Polypropylene (PP), induce a shading
178 effect, limiting energy and substance transfer, ultimately resulting in reduced nutrient and light
179 uptake (Thiagarajan et al., 2021; Wang et al., 2020; Zheng et al., 2021). Microplastic algal
180 interactions may also result in injured cell structure due to impalement, as well as other
181 morphological changes including cell wall thickening and deformed thylakoids (Mao et al.,
182 2018; Zhang et al., 2017). This ultimately results in a reduction of Chlorophyll content and
183 growth (Thiagarajan et al., 2021; Wang et al., 2020; Zheng et al., 2021). Because epiphytes act
184 as a significant trap for microplastics, and microplastics are often found adhered to and/or

185 covering epiphytes, shading and impalement may be more likely to occur than in free floating
186 marine algae (Gerstenbacher et al., 2022).

187 Microplastics further threaten marine algae through the development of biofilms; a thin
188 film of bacteria and algae that sticks to microplastic surfaces, whose formation is caused by a
189 combination of environmental and plastic attributes, including surface roughness, charge,
190 electrostatic interactions, light, and temperature (Rummel et al., 2017, Oberbeckmann et al.,
191 2015). Biofilm formation predisposes microplastic-algae hetero-aggregations; microplastics
192 devoid of biogenic particles do not aggregate, whereas those covered in biofilms can form
193 aggregates, interact with other existing aggregates, and lend to the creation of new aggregates
194 (Michels et al., 2018). These aggregates often cause marine algae to precipitate or be rendered
195 inactive, reducing photosynthesis and growth (Long et al., 2017; Zhang et al., 2017). Biofilms
196 also increase attributes of microplastics, such as charge and roughness, which allows them to
197 vector heavy metals like Ag, Ni, Pb, and Zn (Wang et al., 2016). Marine algae can absorb these
198 metals and have shown decreased growth and chlorophyll content under heavy metal stress
199 (Saleh, 2015; Sheng et al., 2004). Similarly, to other pollutants, microplastics may make heavy
200 metals more locally abundant to epiphytes when they become trapped amongst them.

201 Seagrass epiphytes have thus far been majorly overlooked when considering microplastic
202 ecosystem impacts, though the majority of microplastics on seagrass blades are adhered to
203 epiphytes (Goss et al., 2018; Gerstenbacher et al., 2022; Table 1). Because epiphytes perform
204 vital ecosystem functions by enhancing spatial complexity, above ground biomass, and primary
205 production, identifying potential effects associated with their microplastic contamination is vital
206 when assessing overall ecosystem health (Bologna et al. 1999; Borowitzka et al., 2007).

207 While microplastics are capable of significantly reducing algal growth and photosynthesis, net
208 impacts on seagrasses themselves may be more complex. If epiphyte growth and photosynthesis
209 is decreased due to microplastics, this may benefit the seagrass plant, as epiphytes are direct
210 competitors with seagrasses for light, nutrients, and space (Bulthuis and Woelkerling, 1983;
211 Brodersen et al., 2015; Brodersen et al., 2020). A reduction in epiphytes may lead to increased
212 passive diffusion of CO₂, O₂, and nutrients into the blades, and higher inorganic carbon
213 availability in the leaf microenvironment (Brodersen et al., 2015; Brodersen et al., 2020).
214 However, these benefits can only be achieved if microplastics do not also have negative impacts
215 on seagrasses themselves. While the breadth of literature on this topic is limited, evidence
216 suggests this is unlikely. Microplastics adhered to seagrass blades may act physically like
217 epiphytes themselves by obstructing seagrass cells, creating a shading effect and subsequently
218 decreasing light attenuation and nutrient transfer (Silberstein et al., 1986; Brodersen et al., 2020).
219 From a toxicity perspective, microplastics may make pollutants more locally abundant on and
220 surrounding seagrass blades via desorption (Heinrich & Braunbeck, 2019). While documentation
221 of toxin bioavailability via microplastic desorption does not yet exist for seagrasses, evidence in
222 marine and aquatic algae suggest the pathway's viability (Ge et al., 2021; Yi et al., 2020). Under
223 circumstances where pollutants become bioavailable to plant surfaces, seagrasses will likely have
224 adverse reactions (Adamakis et al., 2021; Ralph & Burchett, 1998; Macinnis-Ng & Ralph, 2002,
225 Zheng et al., 2019). Heavy metals such as Zn and Cu have negative effects on several species of
226 seagrasses' photosynthetic rates, reducing chlorophyll *a* fluorescence in instances of only 1 mg
227 litre⁻¹ of exposure, and causing metabolic and morphologic effects in many instances (Ralph &
228 Burchett, 1998; Macinnis-Ng & Ralph, 2002, Gaeckle et al., 2012; Zheng et al., 2019). The
229 seagrass *Cymodocea nodosa* was negatively impacted by exposure to environmentally relevant

230 quantities of bisphenol A (BPA), where cells lost chlorophyll auto-flourescence and accumulated
231 H₂O₂. Further, polycyclic aromatic hydrocarbons (PAHs) have been found accumulating and
232 inhibiting growth in thylakoid membranes of chloroplasts in the seagrass *Posidonia oceanica*, as
233 well as other aquatic plants (Apostolopoulou et al., 2014; Adamakis et al., 2021).

234 Microplastic adherence to seagrass blades may place not only epiphytic communities, but
235 seagrasses themselves at risk. Microplastics likely have the capacity to physically obstruct plant
236 cells and increase pollutant concentrations in the blade microenvironment, ultimately leading to
237 reduced photosynthesis and growth. Seagrasses are valued highly for services related to growth
238 and photosynthesis, including carbon sequestration and wave attenuation (Nordlund et al., 2018).
239 Because these services are already at risk from many other local and global stressors, new and
240 increasing threats such as microplastics, whose impact severity is largely unknown, will be
241 valuable to study when considering the overall health and stability of seagrasses now and in the
242 future (Duarte, 2002; Unsworth et al, 2018).

243 **Impacts of Microplastics on Seagrass Sediment Communities**

244 Just as microplastics get trapped amongst seagrass blades, they can become accreted into
245 the ecosystem's sediments. In fact, seagrass bed sediments are a significant sink for
246 microplastics compared to bare sediments, due mainly to near-bed turbulent kinetic energy
247 (TKE), the same process which leads to sediment trapping (Huang et al., 2020; Smit et al., 2021)
248 Seagrass blades with adhered microplastics may also shed to the sediments below, contributing
249 to the stock (Huang et al., 2021). Furthermore, aggregations of microplastics and their associated
250 biofilms result in an increase in density, causing rapid sinking into sediments (Leiser et al., 2021;
251 Rummel et al., 2017). Huang et al. (2020) found that microplastics within seagrass meadow
252 sediments were enriched by a factor of up to 2.9 compared to bare sediments. While there is a

253 considerable amount of research detailing adverse impacts of microplastics on benthic
254 organisms, these impacts are not often considered in the context of seagrass ecosystems’
255 significant microplastic enrichment, where effects may be heightened and there is a higher
256 density and diversity of flora and fauna (Huang et al., 2020; Browne et al., 2008; Sussarellu et
257 al., 2016; Duffy, 2006). Increased microplastic density may heighten local sediment pollutant
258 concentrations, as microplastics vector heavy metals, residual monomers, and other pollutants
259 into marine sediments (Li et al., 2020; Cauwenberghe et al., 2015). This could cause higher loads
260 of microplastics and associated toxins in the bodies of benthic organisms like mussels (*Mytilus*
261 *edulis*), oysters (*Crassostrea gigas*), and crustaceans (*Carcinus maenas* & *Nephrops norvegicus*).
262 These species, and many others, have all had negative reactions to relevant quantities of
263 microplastics, including digestive inflammation, reduced fertility, reduced feeding, and trophic
264 transfer (Cauwenberghe et al., 2015; Browne et al., 2008; Sussarellu et al. 2016). Benthic
265 organisms in two seagrass ecosystems in Indonesia show a microplastic contamination level of
266 28.29% and 25%, indicating that high levels of contamination within the meadows correspond to
267 high contamination in benthic organisms (Tahir et al., 2019). Sufficient quantities of
268 microplastics are also capable of altering mean grain size, permeability, and decreasing thermal
269 diffusivity, impacting some organisms and their eggs, which prefer certain sediment sizes,
270 oxygen concentrations, moisture contents, and temperatures (Wang 2016). Harming benthic
271 organisms would negatively impact many aspects of the ecosystem, as they perform key
272 ecosystem functions like water filtration, bioturbation, predation/grazing, and organic matter
273 decomposition (Duffy et al., 2006).

274 Microplastics within sediments may also impact microbial and plant communities,
275 particularly by altering nutrient cycling. Seeley et al., 2020 shows that microplastics prevent

276 nitrification and denitrification in sediment dwelling microbes, while van Weert et al. 2019
277 reveals microplastics experimentally decrease nutrient uptake and shoot to root ratios of
278 sediment rooted macrophytes (Yao et al., 2019). High microplastic densities within seagrass
279 sediments, which will likely increase local toxin concentrations, could also inhibit dinitrogen
280 fixation in the diazotrophic flora on the roots and rhizomes of seagrasses, as seen in the seagrass
281 species *Zostera marina* (Brackup & Capone, 1985). These impacts may alter nitrogen dynamics
282 within the ecosystem, causing a limitation. As many seagrass ecosystems are already limited by
283 nitrogen, any further alteration to this cycle will be vital to understand (Johnson et al., 2006;
284 Powell et al., 1989). Regarding carbon sequestration, Huang et al., 2021 shows that the
285 accumulation of particulate organic carbon (POC) and microplastics within seagrass sediments is
286 linked, as higher POC is correlated to higher microplastic abundance. There is currently no
287 published work done exploring the impacts of this association, nor regarding microplastic's
288 influence on carbon cycling within seagrass sediments. Future research should focus on
289 identifying the capacity of microplastics to influence these processes (Shen et al., 2020).

290 **Microplastic Impact Implications and Gaps**

291 Microplastics retained within seagrass sediments and blades may have the ability to
292 complexly alter ecosystem functions with their chemical and physical attributes. Epiphyte and
293 seagrass photosynthesis and growth, nutrient cycling, and benthic organism health and function
294 could be impacted (Fig. 1). Effects may be exacerbated further by seagrasses' high microplastic
295 trapping efficiency, and dense and diverse flora and fauna (Huang et al., 2019). Currently,
296 microplastic concentrations on Turtle Grass (*Thalassia testudinum*) blades with high epiphytic
297 loads average 4.0 ± 2.1 microfibers and 0.714 ± 0.42 microbeads per blade (Goss et al., 2018).
298 Because microplastic concentrations in the low ppm cause negative impacts to marine algae,

299 there is reason to believe that microplastic concentrations on seagrass blades may already cause
300 adverse effects (Prata et al., 2019). Similarly, Huang et al., 2019 found microplastic
301 concentrations ranging from 80 to 884.5 particles/kg of dry seagrass sediment in China, enriched
302 by factors of 2.1 and 2.9 relative to bare sediments. Experimentally, 500 particles/kg or less
303 cause negative impacts to the survival and growth of benthic organisms (Ziajahromi et al., 2018).
304 It is possible that current microplastic concentrations within seagrass sediments are negatively
305 impacting many aspects of the ecosystem which are vital to its function, including microbial and
306 plant nutrient dynamics, and benthic organism functions.

307 Existing literature on this topic has focused primarily on documenting microplastics
308 within seagrass ecosystems and exploring the mechanisms of their accumulation (Table 1).
309 While this body of work is by no means complete, as it becomes evident that microplastics are
310 pervasive within seagrass meadows with little hope for removal, literature must pivot towards
311 exploring the impacts of this accumulation. Microplastic quantification within seagrass
312 ecosystems is becoming increasingly available for several seagrass species (Table 1). With this,
313 it is possible to set up in vitro experimentation using environmentally relevant microplastic
314 concentrations, to explore their impact on epiphytic and plant photosynthesis and growth,
315 nitrogen and carbon cycling, and benthic organism health. This review informs and motivates
316 such work, so we may understand the effects of microplastics on vital ecosystem members and
317 processes. Only then can we properly assess the health and longevity of seagrasses in the
318 plasticine.

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326 **Author Contributions**

327 **C.M.G:** Conceptualization, Writing: original draft preparation, Funding acquisition,
328 Visualization. **A.C.F:** Conceptualization, Writing: reviewing and editing, Supervision. **A.B.N:**
329 Writing: reviewing and editing, Funding acquisition, Supervision. **R.R:** Writing: reviewing and
330 editing, Funding acquisition, Supervision.

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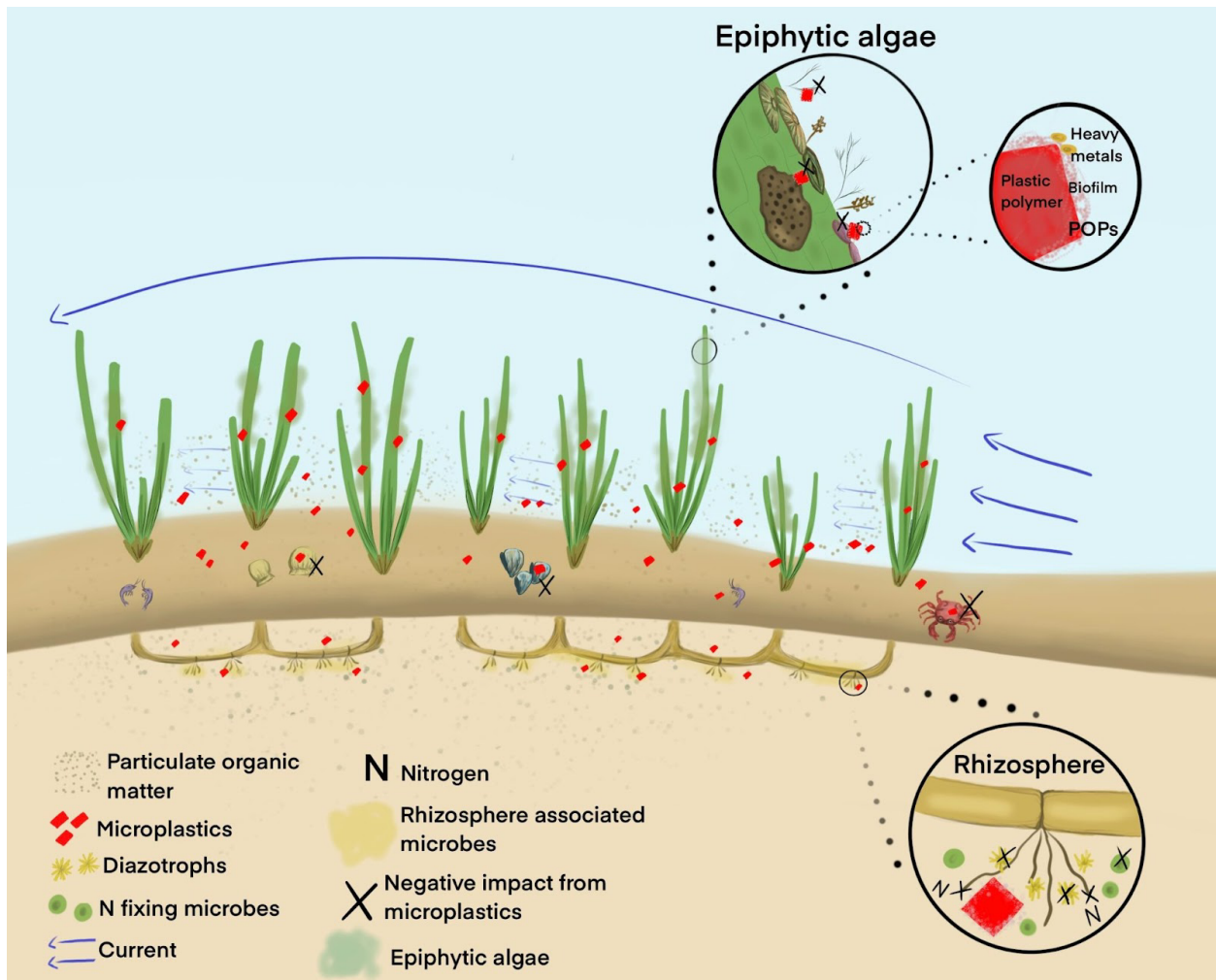
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573 **Figure 1. Conceptual model of microplastics and their sorbed pollutants' impact on**

574 **seagrass communities.** Model demonstrates microplastics sinking within seagrass ecosystems,

575 accumulating both amongst epiphytic algae adhered to seagrass blades and within sediments.

576 Biological members of the ecosystem whose function may be compromised by high microplastic

577 concentrations are delineated by a black X; these include: rhizospheres and associated microbes,

578 Benthic organisms like crustaceans and mollusks, epiphytes, and seagrass plants. Considering

579 seagrasses' considerable microplastic trapping ability, many ecosystem members and processes

580 may be compromised.

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582 Table 1: Comprehensive list of papers regarding microplastic accumulation and/or impacts in
 583 seagrass ecosystems.

Topic	Location	Comments
Seagrass Plants and Associated Epiphytes	Turneffe Atoll, Belize (Goss et al., 2018); Barrang Caddi Island, Indonesia (Datu et al., 2019); Pramuka Island, Indonesia (Pricilla et al., 2019); Deergrass Sound, Orkney, Scotland (Jones et al., 2020); Singapore (Seng et al., 2020); Mallorca Island, Western Mediterranean Sea (Sanchez-Videl et al., 2021); Ria Formosa, Portugal (Cozzolino et al., 2020)	No peer reviewed work regarding microplastics on seagrasses exists for North America and Australia. Most articles to date quantify microplastics on blades, but do not assess impacts to plant and epiphyte biology.
Sediments	Makassar Strait, Indonesia (Tahir et al., 2019); Southern China Sea (Huang et al., 2021); Xincun Bay and Li'an Bay, China (Huang et al., 2020); Deergrass Sound, Orkney, Scotland (Jones et al., 2020); Baltic Sea (Kreitsberg et al., 2021); Southern England & Wales (Unsworth et al., 202); Rameswaram Island, India (Jeyasanta et al., 2020); Ria Formosa, Portugal (Cozzolino et al., 2020); New Hampshire, USA (Cheng et al., 2021)	Most articles focus on accumulation and quantification, but do not assess impacts on biogeochemistry and sediment characteristics.
Benthic organisms	Ria Formosa, Portugal (Cozzolino et al., 2020); Pramuka Island, Indonesia (Pricilla et al., 2019); Deergrass Sound, Orkney, Scotland (Jones et al., 2020); Florida Keys, USA (Plee et al., 2020); Barranglombo Island, Makassar, Indonesia (Sawalman et al., 2021)	There is an extensive body of literature that assesses microplastic accumulation in marine organisms and explores impacts to biology. (Miller et al., 2020). However, few studies are directly associated with seagrass ecosystems.
Mesocosm studies assessing retention and/or impacts	Gullmars Fjord, Sweden (De los Santos et al., 2021); Red Sea (De Smit et al., 2021)	Mesocosm experiments assessing microplastic accumulation and impacts remains the most understudied topic.

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